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3
4 **Breeding on the edge: Estimating sex ratios of loggerhead turtles *Caretta caretta***
5 **at the northern extent of their range.**

6
7 **Project Role:**

8 The majority of my data was collected by 51 sea turtle monitoring programs in
9 each of the 3 states. Most of the data I used was from 2000-2006 for some states and
10 sites data went as far back as the 80s. I received data from each state's natural
11 resources agency, which collates and organizes data every year. I converted data
12 prior to 2000 from SC to electronic format. I spent about 4 months reorganizing and
13 reformatting the data, checking for mistakes and pulling out the information I needed.
14 DuBose Griffin and Matthew Godfrey, both state coordinators, helped me decipher
15 this data.

16 I coordinated most site visits and did all the sand collection. I programmed
17 and put out all the loggers at each site except two in NC. Staff from each site usually
18 helped me with transport onto the beaches. I also helped with spectroradiometer
19 readings in the lab in Edinburgh. Alasdair Mac Arthur converted the readings into
20 albedo values for each sample because I did not have the software or the expertise that
21 he did. My thesis was edited by Brendan, DuBose and Matthew.

22
23 **Problems:**

24 The biggest problem I had was making sense of data that wasn't collected by
25 me. Trying to reformat and reorganize it so everything was consistent took a lot of
26 time and effort. I ran into many obstacles of inconsistency and entry errors.

27 Losing loggers was another problem. It probably would have been better if
28 my supervisor could have come out in the field with me in the beginning. I think I
29 would have labeled them better and been more careful about where I put them with
30 his guidance. I think I underestimated the power of the wind and sea!

31 I also had some issues with time. It seemed all a bit rushed at the end because
32 my fieldwork ended so late and I think I might feel more confident in my work if I
33 had more than a month to write it up.

34
35 **Journal: Marine Ecology Progress Series**

36 I chose this journal because it is the leading journal in its field. It covers all
37 aspects of marine ecology, fundamental and applied. I feel my study contains both
38 fundamental and applied aspects of marine ecology. Studying sex ratios is a
39 fundamental part of understanding the total ecology of sea turtles but can also be
40 applied to help conserve and manage the species.

41

42 **Breeding on the edge: Estimating sex ratios of loggerhead turtles *Caretta caretta* at the**
43 **northern extent of their range**

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54 Running head: Sex ratios of northerly loggerheads

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56 Journal: Marine Ecology Progress Series

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58

59 ABSTRACT

60 Temperature-dependent sex determination (TSD) has been studied in loggerhead sea turtles
61 for a number of decades. Most loggerhead populations have been found to exhibit
62 predominately female biased primary sex ratios. Here we study a sub-population in the North
63 Atlantic which has been thought to have a less skewed sex ratio. Temperature and sand were
64 collected from 51 sites along the southeastern United States in order to explain patterns of sex
65 ratios. Incubation durations were calculated for over 18,000 nests from 2000-2006 and
66 converted to sex ratios. Temperature and therefore incubation duration correlated
67 significantly with latitude. Deviations from this trend were explained by sand albedo. On
68 average sex ratios become gradually less female-biased as latitude increases, with the most
69 northern sites producing male-biased clutches. The overall sex ratio was 53.4% female.
70 Long term trend analysis of sex ratios revealed one site had a significant increase in females
71 from 1980-2006. Northern nesting sites are essential for producing consistent numbers of
72 males in an overall heavily biased population. These males will become increasingly more
73 important if climate change causes southern sites to produce further biased primary sex ratios.
74 Continual protection from harmful anthropogenic influence, especially that which could alter
75 the thermal habitat, must be a high priority for conserving and managing this species.

76

77

78 KEYWORDS: loggerhead turtle, *Caretta caretta*, temperature-dependent sex determination
79 (TSD), incubation duration, latitude, climate change, sand albedo

80

81 **INTRODUCTION**

82 **Temperature-dependent sex determination (TSD)**

83 Sex determination of an individual, and therefore the sex ratio of an entire population, can
84 either be influenced by genetic and/or environmental factors. Temperature is the most
85 studied and most influential environmental factor for differentiating sex in vertebrates
86 (Charnier 1966, Yntema 1979, Yntema & Mrosovsky 1980, Mrosovsky 1988; Ciofi &
87 Swingland 1997, Chevalier et al. 1999, Shine 1999, Godfrey and Mrosovsky 2001, Azuma et
88 al. 2004, Robert et al. 2006). Temperature-dependent sex determination (TSD) often has a
89 specific range of values which produces both sexes at varying percentages and at the pivotal
90 temperature the sex ratio is 1:1. Above and below this range offspring are either all male or
91 all female (Yntema & Mrosovsky 1982, Ciofi & Swingland 1997). Actual patterns and
92 ranges differ for each species: some produce all male at higher temperatures and others all
93 female. These temperatures trigger genes and hormones that cause the differentiation of
94 gonads after an egg has been fertilized (Ciofi & Swingland 1997, Pieau & Dorizzi 2004).
95 The evolution of TSD and its adaptive significance is still not fully understood (Shine 1999,
96 Leimara et al. 2004, Valenzuela 2004).

97

98 **TSD in reptiles**

99 There is much debate about the evolution of TSD in reptiles, even though they have been
100 studied for over 4 decades (Charnier 1966, Yntema 1979, Yntema & Mrosovsky 1980,
101 Mrosovsky 1988, Bull & Charnov 1989, Ciofi & Swingland 1997, Chevalier et al. 1999,
102 Godfrey et al. 1999, Shine 1999, Robert et al. 2006). It is unclear why some reptiles have
103 genetic sex determination (GSD) while others have environmental sex determination (ESD).
104 The type of sex determination is not strongly associated with evolutionary linkage (Janzen &
105 Krenz 2004). Neither seems to be the ancestral trait; they could have evolved multiple times
106 or have different origins (Ciofi & Swingland 1997). Even the pattern of sexual
107 differentiation response to temperature varies within reptile taxa (Ewert et al. 2004, Harlow

108 2004). Nest incubation temperatures determine the sex in TSD reptiles. The temperature
109 inside the nest in turn is dependent on several factors, including: nest material/substrate and
110 its subsequent grain size, porosity, water content and albedo; nest location in regards to both
111 depth and linear space; the number of eggs in a clutch which affects the metabolic heating of
112 the nest; and climatic conditions such as air temperature and rainfall (Janzen 1994, Hays et al.
113 1995, Ackerman 1997, Ciofi & Swingland 1997, Shine 1999, Janzen & Morjan 2001, Hays et
114 al. 2001, Miller et al. 2004).

115

116 **Loggerhead sea turtles**

117 One of the most heavily studied reptiles with TSD is the loggerhead sea turtle, *Caretta*
118 *caretta* (Yntema & Mrosovsky 1980, Yntema & Mrosovsky 1982, Limpus et al. 1983,
119 Maxwell et al. 1988, Mrosovsky 1988, Marcovaldi et al. 1997, Hanson et al., 1998, Godley et
120 al. 2001a b, Mrosovsky et al. 2002, Kaska et al. 2006). Loggerhead clutches have a viable
121 thermal range from about 25-35°C with lower extremes producing all males, upper extremes
122 producing all females and a middle range producing both sexes (Yntema & Mrosovsky 1982,
123 Ackerman 1997). Loggerheads are widely distributed around the world, from tropical to
124 warm temperate waters in the Atlantic, Indian, and Pacific Oceans as well as the
125 Mediterranean and the Caribbean Seas. Despite the large range, loggerheads typically have a
126 pivotal temperature between 28.6 - 29.2 °C (Limpus et al. 1985, Mrosovsky 1988,
127 Marcovaldi et al. 1997) and pivotal incubation duration—the time period giving an even sex
128 ratio—between 59.3 - 61.7 days (Godfrey & Mrosovsky 1997, Marcovaldi et al. 1997).
129 These slight variances in pivotal temperatures and incubation durations might be due to the
130 limitations of direct sex ratio sampling when creating sex ratio models and/or differences in
131 laboratory conditions (Mrosovsky 1988).

132

133

134

135 **Limitations of sex ratio sampling**

136 Classifying sex in young sea turtles is often difficult because they do not display external
137 dimorphic sexual characteristics nor do they have dimorphic sex chromosomes (Bull 1980,
138 Ewert 1991). Therefore sex is usually assigned by microscopic visualization of the gonads, a
139 technique that unfortunately results in the death of the hatchling. Because hatchlings must be
140 killed, direct sex ratio sampling is limited because loggerheads are a protected species in
141 much of their range. Often the studies relating sex ratio to temperature only include a few
142 clutches (Yntema & Mrosovsky 1982, Limpus 1983, Mrosovsky 1988).

143

144 However these models make it possible to estimate sex ratios indirectly on a large scale
145 through sand temperatures (Mrosovsky & Provancha 1992, Hanson et al.1998, Godley et al.
146 2001b) and even incubation durations (Marcovaldi et al. 1997, Mrosovsky et al. 1999,
147 Godley et al. 2001a). Estimating sex ratio indirectly is less accurate for individual hatchlings,
148 but allows for greater spatial and temporal extent (Mrosovsky et al. 1999). Using incubation
149 durations especially allows for long-term sex ratio estimates because archived nest data can
150 be used. Still, temperatures and incubations from multiple years from only one beach may
151 not give an accurate estimate of an entire population's sex ratio. Information from a variety
152 of nesting beaches as well as long term sex ratio information is preferred as incubating
153 temperatures can change dramatically from beach to beach and from year to year (Mrosovsky
154 et al 1984, Godfrey & Mrosovsky 1999).

155

156 **Loggerheads in the U.S.A.**

157 Loggerheads in the United States nest along the southern Atlantic coast and are the second
158 largest population of loggerheads in the world (Magnuson et al. 1990, TEWG 1998, TEWG
159 2000). The majority of the population nests in Florida, however their range extends from
160 Texas through Georgia, the Carolinas, and even Virginia, spanning a multitude of beaches
161 and temperatures. Most of the studies in the past have focused on the turtles nesting in

162 Florida since it holds the majority of the population. These studies have shown that turtles in
163 Florida are up to 90% female (Mrosovsky & Provanha 1992, Hanson et al. 1998). This is
164 not unique as other populations in the Mediterranean and in Brazil are also predominately
165 female (Marcovaldi et al. 1997, Godley et al. 2001a, b). Protecting male producing beaches
166 is critical for the population's survival (Baptistotte et al. 1999). Initial studies covering some
167 of the most northerly beaches in this population's range, which have cooler temperatures,
168 suggest they have a near balanced sex ratio (Mrosovsky et al. 1984, Mrosovsky 1988,
169 Hawkes et al. 2007). This may indicate that the northern sites have extreme conservation
170 value by virtue of their consistent male production, and thus require special attention. Studies
171 of Brazilian loggerheads have shown that cooler beaches, though smaller in nest numbers, are
172 important for producing males into the population there (Marcovaldi et al. 1997, Baptistotte
173 et al. 1999, Mrosovsky et al. 1999). However, there has not been a large scale study
174 covering the entire extent of northern loggerhead nest sites in the U.S.A.

175

176 **Climate change**

177 Climate change could have a significant impact on a species that is innately connected with
178 temperatures of its habitat (Janzen 1994, Davenport 1997, Fish et al 2005, Salinger 2005,
179 Hawkes et al. 2007). With increasing temperatures, loggerhead populations that are already
180 skewed, may become entirely female. Some clutches may even extend above thermal limits
181 for successful incubation (Davenport 1997, Matsuzawa et al. 2002). Thus it is important to
182 understand historical and present population trends for the entire population in the face of
183 possible large scale changes to their environment.

184

185 In this study we attempt to examine the sex ratios of the entire northern nesting population of
186 loggerhead turtles in the southeastern United States. We will use incubation duration to
187 estimate hatchling sex ratios using long-term data from monitored nesting beaches in

188 Virginia, North Carolina and South Carolina. It is expected that there will be temporal,
189 latitudinal and sand albedo effects on sex ratio.
190

191 **MATERIALS AND METHODS**

192 **Nesting data**

193 Nesting data were collected by a range of government and non-government agencies over the
194 past few decades by daily nest monitoring programs at 51 beaches from Virginia, North
195 Carolina and South Carolina in the southeastern United States. Sites spanned from 32° 09' N
196 to 37° 56' N latitude. Nesting data were collated and stored by each state's natural resources
197 agency. Incubation duration and hatch success were extracted from these datasets from years
198 which were thought to have complete coverage of each state. When compiling records from
199 all three states, only data from 2000-2006 were used. Incubation duration was recorded as
200 the number of days between egg laying and first emergence of the hatchlings. Because
201 monitoring was mostly carried out in the morning, the date laid and emergence date were
202 scored as the date of the morning after these events occurred the previous night. Several days
203 after the first hatchling emergence, nests were excavated to count the number of egg shells
204 and any eggs or hatchlings, dead or alive, which were left in the nest. Hatch success was
205 calculated as the total number of empty egg shells divided by the total number of eggs laid
206 subtracting any dead hatchlings found for each year for each site. Hatch success was only
207 calculated for nests with clutch and emergence information available. Sex ratios were
208 calculated using published models relating percent female to incubation duration for
209 loggerheads in the U.S.A. (Godley et al. 2001a, Godfrey & Mrosovsky 1997). Sex ratios
210 were only calculated for nests with incubation duration and did not take into account hatch
211 success.

212

213 **Sand temperature**

214 Twelve of the fifty-one sites were monitored for sand temperature during the 2007 nesting
215 season. Three data loggers were placed at each site. In Virginia and South Carolina Tiny
216 Talk II Gemini data loggers (Chinchester, UK) were used and in North Carolina Hobo data
217 loggers (Hobo u-pendants, Onset computers, Massachusetts, U.S.A) were used. Data loggers

218 registered temperature measurements every four hours. Data loggers were placed in the sand
219 45 cm deep which is mean nest depth for loggerhead turtles in this area (Hawkes et al. 2007).
220
221 Sand temperatures were averaged to give daily mean temperatures for each beach. Some
222 loggers were lost, therefore some beaches did not have all three data sets to average. Because
223 of limitations in field time, temperatures were only collected over the first two months of the
224 four month nesting season. When temperatures were compared only values from overlapping
225 days were used which was from 27 May -27 June and this period constitutes the first third of
226 the nesting season. To give an indication of relative thermal environment at these beaches
227 deviance from the median temperature value (25.8°C) was calculated for each site.

228

229 **Sand absorption measurements**

230 Sand was collected from the surface of the beach of most sites along the coast and measured
231 for broad-band absorption (α) using a spectroradiometer (FieldSpec FR, Analytical Spectral
232 Devices, Boulder, CO, U.S.A.) which measures the reflectance in 1 nm increments between
233 350 and 2500 nm of light. We repeated the exact same procedures as in Hays et al. (2001) to
234 obtain albedo (ρ) or reflectance values for each sand sample in the lab. Albedo is the
235 proportion of incoming solar radiation that is reflected back into the atmosphere and is the
236 complementary value to absorption. When albedo is high, absorption is low leading to less
237 heating of the surface from solar energy. Absorption values were obtained by subtracting the
238 albedo values from 100:

$$239 \alpha_{350-2500}(\%) = 100 - \rho_{350-2500}$$

240

241

242 **RESULTS**

243 **Latitudinal Comparisons**

244 Latitudinal comparisons were made for 47 beaches which had consistent monitoring from
245 2000-2006 (Figure 1a). Number of nests varied greatly from beach to beach because of
246 differences in length and latitude. Comparing annual number of nests per km shows a large
247 range of nesting densities from less than 1 nest km⁻¹ yr⁻¹ to over 66 nests km⁻¹ yr⁻¹. Larger
248 nesting areas are located in South Carolina, followed by North Carolina, with very low
249 nesting densities in Virginia (Figure 1b).

250

251 The mean incubation duration for the entire northerly population from 2000-2006 was 61.2 ±
252 5.6 days (range= 54.7 - 90 days, *n* = 47) (Figure 1c). Incubation duration generally increased
253 as latitude increased. The mean site specific sex ratio was 53.4% ± 0.146 female (range= 0 -
254 77%, *n* = 47) (Figure 1d) but this proportion decreased with increasing latitude. Mean hatch
255 success rate for the entire study area was 71.1% ± 0.129 (range= 16 - 93%, *n* = 47) (Figure
256 1e). There seems to be a peak in hatch success in the middle latitudes, in southern North
257 Carolina and northern South Carolina. Hatch success was lowest in Virginia.

258

259 **Range of incubation durations and sex ratio**

260 In Virginia incubation durations ranged from 55-97 days and 0-75% female (Figures 2a and
261 b). In North Carolina incubation durations ranged from 43-99 days and 0-95% female
262 (Figures 2c and d). In South Carolina incubation durations ranged from 43-90 days and 0-
263 95% female (Figures 2e and f).

264

265 Incubation durations varied considerably by state (Figures 2a, c and e). The majority of nests
266 in Virginia and South Carolina were above and below the pivotal incubation duration
267 respectively (Figures 2a and e). However in North Carolina nests were split almost evenly
268 above and below the pivotal incubation duration (Figure 2c). A complementary pattern was

269 seen in the sex ratios of each state (Figures 2b, d, and f). Given the South Carolina bias in
270 magnitude of nests, the entire population from 2000-2006 had incubation durations below the
271 pivotal incubation period thus making most nests predominantly female (Figures 2g and h).

272

273 **Total nests and hatchlings**

274 There were approximately 18,000 loggerhead turtle nests monitored in Virginia, North
275 Carolina and South Carolina from 2000-2006, which is ~ 2575 nests annually. Virginia,
276 North Carolina and South Carolina received 0.2%, 26% and 73.8% of the total number of
277 nests laid in this region respectively. There were ~1.2 million hatchlings that emerged from
278 these nests. Average sex ratio was 53% female and 47% male. Nests laid in Virginia and the
279 Carolinas from 2000-2006 account for ~ 4% of the total population in the U.S.A. (Table 1).

280

281 **Latitude and sand absorptance models**

282 Sand absorptance ($\alpha_{350-2500}$) values ranged from 45.75 - 67.54% (Figure 3). The average raw
283 absorptance value was $56.47\% \pm 0.05$ ($n = 57$). Using a general linear model (GLM)-
284 ANOVA to test for significance, temperatures were highly correlated with latitude ($F =$
285 $45.78, p < 0.001$) and slightly less correlated with absorption ($F = 5.94, p < 0.05$). The
286 interaction between latitude and absorption did not have a significant affect on temperature (F
287 $= 0.09, p > 0.05$) (Figures 4a and c). GLM-ANOVA analysis also showed correlation
288 between incubation durations and latitude ($F = 13.91, p < 0.001$) and absorption ($F = 7.67, p$
289 < 0.01). The combined interaction between latitude and absorption also had a significant
290 affect on incubation duration ($F = 8.12, p < 0.01$) (Figures 4b and d).

291

292 **Long term trends in sex ratio**

293 At a number of sites long-term datasets were available going back a few decades which were
294 utilized to examine any temporal trends in sex ratio. There no was trend in sex ratio for most
295 sites except for Bald Head Island (Figure 5). In Bald Head Island, the percentage of females

296 is increasing annually (regression: $F_{1,24} = 17.6, p < 0.001, R^2 = 0.44$). Annual fluctuations
297 can be seen in all sites. There seems to be more variation in annual sex ratio in northern
298 sites—Virginia and North Carolina—than sites in South Carolina. There also seems to be an
299 increasing trend from 2000-2006 for most sites.

300

301

302

303 **DISCUSSION**

304 **Latitudinal comparisons**

305 Incubation durations and resulting sex ratios followed the predicted patterns of variance with
306 latitude. On average, incubation durations were shorter and the percentage of females greater
307 in lower latitudes compared with higher ones. This pattern is consistent with previous
308 research which showed a higher percentage of males in northern sites compared with sites in
309 Florida (Mrosovsky et al. 1984, Mrosovsky and Provanha 1992, Hawkes et al. 2007).

310

311 The pattern in hatch success is interesting because it does not follow other latitudinal trends.
312 The peak in the mid-latitudes may mean optimal temperatures and conditions are met there.
313 Coupled with sex ratio, high hatch success in the mid-latitudes may mean more hatchlings are
314 being produced from nests with even sex ratio. Varying numbers of nests could explain the
315 differences between the states. In Virginia, nests numbers are so low that a few unhatched
316 nests could skew the entire average. In South Carolina, there are so many nests, not all can
317 be monitored on an individual level and many are lost to over wash and predation. Also
318 South Carolina has higher temperatures than the other states and higher temperatures have
319 been shown to increase turtle mortality (Yntema & Mrosovsky 1980). Hatch success is also a
320 challenging measure to consistently evaluate across various sites and years. In addition, this
321 study did not include nests for which clutch size was unknown and then were completely
322 destroyed before inventory could occur. This means that hatch success values given here
323 may be slightly higher than the true values. Also, predation by foxes and raccoons may be a
324 serious issue on some beaches which can further distort hatch success.

325

326 Incubation durations and hatch success can also be greatly affected by conservation efforts.
327 Each state has its own regulations for turtle nest monitoring. In South Carolina, for example,
328 eggs can be lost due to probing when initially searching for the nest. This practice is not
329 carried out in Virginia and North Carolina. Also many sites relocate more nests than others

330 due to erosion or interference with human activities. This can skew hatch success and sex
331 ratio because nests are usually moved higher up the beach which can increase the
332 temperatures inside the nest. In some years, as many as 60% of all nests have been relocated
333 in North Carolina and South Carolina. With such large numbers of nests being moved from
334 their original positions, it is likely to change incubation durations but at the same time may
335 ensure more successful nests.

336

337 **Range of Incubation Durations and Sex Ratios**

338 Ranges of incubation durations and sex ratios for each state were interesting because they
339 showed that all states had at least some nests which produced all males (Figure 2). Some of
340 the highest and lowest incubation durations may not be entirely accurate because they fall
341 outside the normal range of incubation. Even with North Carolina and Virginia producing
342 non-female biased nests, the overall range of nests are predominately female biased because
343 South Carolina accounts for over two-thirds of the total. Although most nests were
344 predominately female, the overall mean sex ratio was relatively even. This mean value is
345 based on averages by site so it does not take into account how many nests are at each site.
346 The ranges of sex ratios shown in Figure 2 are a more accurate description of what is
347 happening with nests in Virginia and the Carolinas.

348

349 **Total number of nests**

350 Means calculated for each site do not include incubation and hatch success data for every nest
351 observed. Beach patrols occasionally miss some nests both when eggs are laid and hatching.
352 Some sites do not have the capacity to inventory every single nest. Additionally, storms can
353 wipe out a multitude of nests before any inventory can be done. Also, total number of nests
354 and hatchlings shown in this study do not represent every nest and hatchling produced in
355 these areas. There are still some beaches that are not monitored for sea turtle nesting or have
356 not been monitored long enough to include in this study. Most of the missing sites are remote

357 barrier islands which make daily monitoring difficult. Other sites are not monitored simply
358 because turtle activity is too low to warrant daily beach patrols. South Carolina has the most
359 missing number of nests. Based on aerial surveys on average, one-third of nests in South
360 Carolina are not monitored from the ground (DB Griffin, pers. comm.). This is a large
361 number of nests for which incubation duration data is missing which in turn could affect the
362 overall sex ratios calculated here.

363

364 Virginia, North Carolina and South Carolina make up a small portion of the of 65,040
365 loggerhead nests produced on average each year between 2000-2006 in the United States
366 (Florida Fish and Wildlife Conservation Commission 2007, M Dodd, pers. comm.).

367 According to this study, these states account for about 4% of the North Atlantic loggerheads.

368 This number may be the result of a gradual decline of the northern population. Aerial
369 surveys flown by the South Carolina Department of Natural Resources have shown a 3%
370 annual decline (1980 –2007) in nesting numbers for South Carolina (DB Griffin, pers.

371 comm.). However this decline is not confirmed in other states. But if numbers of turtles
372 from these sites are declining, it is further reason to ensure their protection.

373

374 Although the amount of males being produced at these sites does not make up for the heavily
375 female-biased sites in Florida these sites are very important for the future survival of this
376 population. Climate change may cause total feminization in sites in Florida which are

377 already predominately female. Northerly sites may become the only places in this

378 population's habitat range which are able to produce males. Additionally southern turtles
379 may begin to shift their nesting habits north to these sites (Janzen 1994, Davenport 1997,

380 Hays et al. 2003) or they may change their time of year for nesting. Although they may have

381 low nest numbers now, northern loggerhead nest sites may become the primary nesting areas
382 in the future.

383

384 **Latitude and sand absorption models**

385 Latitude is the primary driver of temperatures and therefore incubation durations. Sand
386 absorption describes the deviance from the expected latitudinal trend. Our sand absorptance
387 values were similar to those sands measured in Hays et al. (2001). There was a temporal
388 mismatch of incubation durations and albedo values which might explain why the combined
389 effect of latitude and absorption on incubation was not more influential than latitude alone.
390 Ideally incubation durations from 2007 should have been compared with sand albedo from
391 2007. Although it may seem like an outlier, significance was not greatly influenced by the
392 long incubation duration from Assateague Island (90 days) because when tested without it,
393 latitude still had a significant affect on incubation.

394

395 Significance values were greater for comparisons with incubation duration than temperature
396 because of a type II statistical error in sample size. If temperature was measured at more
397 sites, the interaction between latitude and absorption would have probably been significant.

398 The latitudinal and absorption trends in temperature are consistent with previous research
399 testing these interactions (Mrosovsky et al. 1984, Mrosovsky et al. 1988, Mrosovsky and
400 Provancha 1992, Hays et al. 1995, Hays et al. 2001, Hawkes et al. 2007). Latitudinal trends in
401 incubation duration for loggerheads in the United States mirror trends found in Brazil
402 (Marcovaldi et al. 1997). Incubation durations have not been compared with absorption as of
403 yet in other studies but should be considered in future research on sex ratios considering its
404 significant effect here.

405

406 **Long Term trends in Sex Ratio**

407 The absence of a temporal trend in sex ratio for most sites may mean that no significant
408 warming has occurred in the study area so far. Not all regions on the earth are experiencing
409 the same amount of warming (IPCC 2001). It is predicted that extreme latitudes will bear the
410 brunt of climate change, but since the sites from this study are in neither of those ranges it is

411 possible temperatures may not significantly rise here for a long time (Hawkes et al. 2007).
412 However it is possible data from some sites did not go back far enough to notice a trend.
413 Bald Head Island, which had a significant trend, had a data set going back 16 years. This site
414 may be an example of what could be seen at other sites if more data was available. Inter-
415 annual differences in temperature may be the source for the differences in sex ratio from year
416 to year however temperature is obviously not the only factor influencing sex. If that were the
417 case, years of high temperatures would produce relative high percentages of females at all
418 sites. There are clearly site-specific factors which influence the nests at each beach. Then
419 again, close examination of the last 6 years of sex ratios at these sites does show an increase
420 in percent females for 5 out of 6 of the sites. It is not a significant trend yet, but may become
421 one if warming increases over the next few years or decades.

422

423 **Study limitations**

424 There are several limitations in this study that could have affected some of the results. Every
425 effort was made to standardize data from all three states. Hatch success and incubation
426 duration calculations are standardized in all states, but only for recent years and that does not
427 mean there were no mistakes made when data were entered. Because this study spanned such
428 a large area and time period, data collection and inventory was a collaborative effort and
429 mistakes were obviously inevitable. For example incubation duration can easily be
430 calculated wrong if the date of first emergence is confused with the date of last emergence or
431 if the evening's date is recorded instead of the date observed. However, we assumed that
432 these errors would occur randomly and thus not greatly bias the results in any particular
433 direction.

434

435 Sand temperatures were only monitored over the first third of the nesting period due to field
436 time constraints. A greater correlation between sand temperature and latitude as well as
437 albedo would have probably been observed if temperatures were monitored over the entire

438 nesting period. Also some sites did not have three data sets of sand temperatures because
439 loggers were lost. On large beaches with varying sand type and usage, one sample of sand
440 temperatures may not be an accurate depiction of the entire beach. Additionally multiple
441 sand samples from each site could have been collected to capture the entire sand albedo
442 spectrum of these beaches. Although other studies have shown little within beach variation
443 in sand albedo (Hays et al. 2001) renourishment and disposal of sand in the United States
444 happens repeatedly along many public beaches and should be considered when collecting
445 sand for sand spectral analysis.

446

447 **Beach alterations**

448 Many alterations to a beach can be made which can change its thermal properties including
449 beach renourishment and disposal but also beachfront development, and/or deforestation.
450 Beachfront development often causes coastal squeeze. This leaves less room on the beach for
451 nests and usually shifts nest further towards the sea which is usually detrimental to hatch
452 success (Carthy et al 2003). High-rise development may also alter the thermal characteristics
453 of a beach by artificially shading certain areas (Mrosovsky et al 1995).

454

455 Beach renourishment occurs when homeowners want to protect their property from being
456 destroyed by storm surge and erosion—which subsequently will increase with warming
457 temperatures and rising sea levels. They refill beaches which have become narrow, in hopes
458 that their properties will be better protected. The sand is usually dredged from the ocean
459 floor and so it can be different from the original sand on the beach. Most ocean bottom sand
460 is full of shells, coarser, and in some cases darker (Ackerman 1997). This could cause
461 warmer incubation temperatures and therefore make northern beaches produce more females
462 than normal. More research is needed testing differences in sand albedo between natural and
463 renourished sand. Deforestation along beach edges could also cause warmer incubation
464 temperatures by removing the shaded areas that allow for cooler temperatures at the top of

465 the beach (Kamel & Mrosovsky 2006). Beach alterations should therefore be kept to a
466 minimum to ensure they do not amplify the affects of global warming.

467

468 **Conclusions**

469 Our study shows that many of the northern nesting sites for North American loggerheads
470 produce a considerable amount of male hatchlings even though these sites only make up a
471 small portion of the total population. The consistent production of some males is important
472 for a loggerhead population which is overall heavily female biased (Mrosovsky & Provanča
473 1992, Marcovaldi et al. 1997, Hanson et al. 1998, Baptistotte et al. 1999, Mrosovsky et al.
474 1999, Godley et al. 2001a b). With a rise in global temperatures, these sites will only become
475 more important for the production of males in the population.

476

477 Our study is useful in that it encompasses multiple sites and multiple years over a continuous
478 latitudinal spread. More studies in the future should attempt to cover such large-scale data
479 sets because they are more accurate in determining the status of a population as a whole. It
480 is clear that northerly nesting sites for these loggerheads should continue to be protected and
481 are of considerable importance for the existence of the population. If there are alterations to
482 these northern sites which can change their thermal properties, these sites may become less
483 capable of ensuring the population's success.

484

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497 LITERATURE CITED

- 498 Ackerman RA (1997) The nest environment and embryonic development of sea
499 turtles. In: Lutz PL, Musick JA (eds) The biology of sea turtles, CRC Press,
500 Boca Raton, p 83–106
- 501 Azuma T, Takeda K, Doi T, Muto K, Akutsu M, Sawada M, Adachi S (2004) The
502 influence of temperature on sex determination in sockeye salmon. *Aquaculture*
503 234: 461–473
- 504 Baptistotte C, Scalfoni JT, Mrosovsky N (1999) Male-producing thermal ecology of a
505 southern loggerhead turtle nesting beach in Brazil: implications for
506 conservation. *Anim Conserv* 2: 9-13
- 507 Bull JJ (1980) Sex determination in reptiles. *Q Rev Biol* 55:3–21.
- 508 Bull JJ, Charnov EL (1989) Enigmatic reptilian sex ratios. *Evolution* 43:1561-1566
- 509 Carthy RR, Foley AM, Matsuzawa Y (2003) Incubation environment of loggerhead
510 turtle nests: effects on hatching success and hatchling characteristics. In:
511 Bolten AB, Witherinton BE (eds) *Loggerhead sea turtles*. Smithsonian Institute,
512 Washington, p 144-153
- 513 Charnier M (1966) Action de la temperature sur la sex-ratio chez l'embryon *d'Agama*
514 *agama* (Agamidae, Lacertilien). *C R Seances Soc Biol Fil* 160: 620–622
- 515 Chevalier J, Godfrey MH, Girondot M (1999) Significant difference of temperature-
516 dependent sex determination between French Guiana (Atlantic) and Playa
517 Grande (Costa-Rica, Pacific) leatherbacks (*Dermochelys coriacea*). *Ann Sci*
518 *Nat* 20: 147-152
- 519 Ciofi C, Swingland IR (1997) Environmental sex determination in reptiles. *Appl*
520 *Anim Behav Sci* 51: 251-265
- 521 Davenport J (1997) Temperature and the life-history strategies of sea turtles. *J Therm*
522 *Bio* 22: 479 488
- 523 Ewert MA, Nelson CE (1991) Sex determination in turtles: diverse patterns and some

524 possible adaptive values. *Copeia* 1991: 50–69.

525 Ewert MA, Etechberger CR, Nelson CE (2004) Turtle sex-determining modes and
526 TSD patterns, and some TSD pattern correlates. In: Valenzuela N, Lance VA
527 (eds) *Temperature-dependent sex determination in vertebrates*. Smithsonian
528 Institute, Washington DC, p 21-41

529 Fish MR, Côté IM, Gill JA, Jones AP, Renshoff S, Watkinson AR (2005) Predicting
530 the impact of sea-level rise on Caribbean sea turtle nesting habitat. *Conserv*
531 *Biol* 19: 482-491

532 Florida Fish and Wildlife Conservation Commission (2007) Statewide nesting beach
533 survey program: nesting data 1990-2006. Accessed 29 Aug.
534 [http://research.myfwc.com/engine/download_redirection_process.asp?file=](http://research.myfwc.com/engine/download_redirection_process.asp?file=Loggerhead_Nesting_Data__1990-2006.pdf&objid=2411&dltype=article)
535 [Loggerhead_Nesting_Data__1990-2006.pdf&objid=2411&dltype=article](http://research.myfwc.com/engine/download_redirection_process.asp?file=Loggerhead_Nesting_Data__1990-2006.pdf&objid=2411&dltype=article)

536 Godley BJ, Broderick AC, Mrosovsky N (2001a) Estimating hatchling sex ratios of
537 loggerhead turtles in Cyprus from incubation durations. *Mar Ecol Prog Ser*
538 210: 195–201

539 Godley BJ, Broderick AC, Mrosovsky N (2001b) Thermal conditions in nests of
540 loggerhead turtles: further evidence suggesting female skewed sex ratios of
541 hatchling production in the Mediterranean. *J Exp Mar Biol Ecol* 263: 45-63

542 Godfrey MH, Mrosovsky N (1997) Estimating the time between hatchling of sea
543 turtles and their emergence from the nest. *Chel Conserv Biol* 2: 581-585

544 Godfrey MH, Mrosovsky N (1999) Estimating hatchling sex ratios. In: Eckert KL,
545 Bjørndal K, Abreu-Grobois FA, Donnelly M (eds) *Research and Management*
546 *Techniques for the Conservation of Sea Turtles*. IUCN/SSC Marine Turtle
547 Specialist Group Publication 4, Blanchard, Pennsylvania, p 136–138

548 Godfrey MH, Mrosovsky N (2001) Relative importance of thermal and nonthermal
549 factors on the incubation period of sea turtle eggs. *Chel Conserv Biol* 4: 217–
550 218

551 Godfrey MH, D'Amato AF, Marcovaldi MA, Mrsovsky N (1999) Pivotal temperature
552 and predicted sex ratios for hatchling hawksbill turtles from Brazil. *Can J Zool*
553 77: 1465-1473

554 Hanson J, Wibbels T, Martin RE (1998) Predicted female bias in sex ratios of
555 hatchling loggerhead sea turtles from a Florida nesting beach. *Can J Zool* 76:
556 1850-1861

557 Harlow PS (2004) TSD in lizards. In: Valenzuela N, Lance VA (eds) *Temperature*
558 *dependent sex determination in vertebrates*. Smithsonian Institute, Washington
559 DC, 42-52

560 Hays GC, Adams CR, Mortimer JA, Speakman JR (1995) Inter- and intra-beach
561 thermal variation for green turtle nests on Ascension Island, South Atlantic. *J*
562 *Mar Biol Ass UK* 75: 405-411

563 Hays GC, Ashworth JS, Barnsley MJ, Broderick AC, Emery DR, Godley BJ,
564 Henwood A Jones EL (2001) The importance of sand albedo for the thermal
565 conditions on sea turtle nesting beaches. *Oikos* 93: 87-94.

566 Hays GC, Broderick AC, Glen F, Godley BJ (2003) Climate change and sea turtles: a
567 150 year reconstruction of incubation temperatures at a major marine turtle
568 rookery. *Glob Change Biol* 9: 642-646

569 Hawkes LA, Broderick AC, Godfrey MH, Godley BJ (2007) Investigating the
570 potential impacts of climate change on marine turtles. *Global Change Biol* 13:
571 1-10

572 Intergovernmental Panel on Climate Change (2001) *Climate Change 2001: The*
573 *Scientific Basis*. In: Houghton JT, Ding Y, Griggs DJ, Noguer M, van der
574 Linden PJ, Xiaosu D (eds) *Contribution of Working Group I to the Third*
575 *Assessment Report of the IPCC*. Cambridge University Press, UK. Accessed
576 29 Jul. <http://www.ipcc.ch/pub/reports.htm>.

577 Janzen FJ (1994) Climate change and temperature-dependent sex determination in

578 reptiles. Proc Natl Acad Sci USA 91: 7487–7490

579 Janzen FJ, Morjan C (2001) Repeatability of microenvironment-specific nesting
580 behaviour in a turtle with environmental sex determination. Anim Behav 62:
581 73–82

582 Janzen FJ, Krenz JG (2004) Phylogenetics: which was first, TSD or GSD? In:
583 Valenzuela N, Lance VA (eds) Temperature-dependent sex determination in
584 vertebrates. Smithsonian Books, Washington, DC, 121-130

585 Kamel SJ, Mrosovsky N (2006) Inter-seasonal maintenance of individual nest site
586 preferences in hawksbill sea turtles. Ecology 87: 2947–2952

587 Kaska Y, Ilgaz Ç, Özdemir A, Başkale E, Türkozan O, Baran İ, Stachowitsch M
588 (2006) Sex ratio estimations of loggerhead sea turtle hatchlings by histological
589 examination and nest temperatures at Fethiye beach, Turkey.
590 Naturwissenschaften 93: 1432-1904

591 Leimara O, Van Dooren TJM, Hammerstein P (2004) Adaptation and constraint in the
592 evolution of environmental sex determination. J Theor Biol 227: 561–570

593 Limpus CJ, Reed PC, Miller JD (1985) Temperature dependent sex determination in
594 Queensland sea turtles: intraspecific variation in *Caretta caretta*. In: Grigg G,
595 Shine R, Ehmann H (eds) Biology of Australasian Frogs and Reptiles. Royal
596 Zoological Society, New South Wales, p 343-351

597 Magnuson JJ, Bjorndal KA, Dupaul WD, Graham GL, Owens DW, Peterson CH,
598 Pritchard PCH, Richardson JI, Sual GE, West CW (1990) Decline of the sea
599 turtles. National Academy Press, Washington, D.C.

600 Marcovaldi MA, Godfrey MH, Mrosovsky N (1997) Estimating sex ratios of
601 loggerhead turtles in Brazil from pivotal incubation durations. Can J Zool 75:
602 755-770

603 Matsuzawa Y, Sato K, Sakamoto W, Bjorndal KA (2002) Seasonal fluctuations in

604 sand temperature: effects on the incubation period and mortality of loggerhead
605 sea turtle (*Caretta caretta*) pre-emergent hatchlings in Minabe, Japan. Mar
606 Biol 140: 639–646

607 Maxwell JA, Motara MA, Frank GH (1988) Microenvironmental study of the effect
608 of temperature in the sex ratios of the loggerhead turtle, *Caretta caretta*, from
609 Tongaland, Natal. S Afr J Zool 23:342–350

610 Miller D, Summers J, Silber S (2004) Environmental versus genetic sex
611 determination: a possible factor in dinosaur extinction? Fertil and Steril 81:
612 954-964

613 Mrosovsky N (1988) Pivotal temperatures for loggerhead turtles (*Caretta caretta*)
614 from northern and southern nesting beaches. Can J Zool 66: 661-669

615 Mrosovsky N, Provancha J (1992) Sex ratio of hatchling loggerhead sea turtles: data
616 and estimates from a 5-year study. Can J Zool 70: 530–538

617 Mrosovsky N, Hopkins-Murphy SR, Richardson JI (1984) Sex ratio of sea turtles:
618 seasonal changes. Science 225: 739 – 741

619 Mrosovsky N, Lavin C, Godfrey MH (1995) Thermal effects of condominiums on a
620 turtle beach in Florida. Biol Conserv 74: 151-156

621 Mrosovsky N, Baptistotte C, Godfrey MH (1999) Validation of incubation duration as
622 an index of the sex ratio of hatchling sea turtles. Can J Zool 77: 831-835

623 Mrosovsky N, Kamel S, Rees AF, Margaritoulis D (2002) Pivotal temperature for
624 loggerhead turtles (*Caretta caretta*) from Kyparissia Bay, Greece. Can J Zool
625 80: 2118–2124

626 Nelson NJ, Thompson MB, Pledger S, Keall SN, Daugherty CH (2004) Do TSD, sex
627 ratios, and nest characteristics influence the vulnerability of tuatara to global
628 warming? Int Cong Ser 1275: 250–257

629 Parmesan C, Yohe G (2003) A globally coherent fingerprint of climate change
630 impacts across natural systems. Nature 421: 37-42

631 Pieau C, Dorizzi M (2004) Oestrogens and temperature-dependent sex determination
632 in reptiles: all is in the gonads. *J Endocrinol* 181:367-377

633 Robert KA, Thompson MB, Seebacher F (2006) Thermal biology of a viviparous
634 lizard with temperature-dependant sex determination. *J Therm Biol* 31: 292–
635 301

636 Salinger M (2005) Climate variability and change: past, present and future – an
637 overview. *Clim Change* 70: 9-29

638 Shine R (1999) Why is sex determined by nest temperature in many reptiles? *Trends*
639 *Ecol Evol* 14: 186-189

640 TEWG (1998) An assessment of kemp's ridley (*Lepidochelys kempii*) and loggerhead
641 (*Caretta caretta*) sea turtle populations in the western North Atlantic. NOAA
642 Technical Memorandum NMFS-SEFSC-409

643 TEWG (2000) Assessment update of kemp's ridley and loggerhead sea turtle
644 populations in the western North Atlantic. NOAA Technical Memorandum
645 NMFS-SEFSC-444

646 Valenzuela N (2004) Evolution and maintenance of TSD. In: Valenzuela N, Lance
647 VA (eds) Temperature dependent sex determination in vertebrates.
648 Smithsonian Institute, Washington DC, 131-147

649 Yntema CL (1979) Temperature levels and periods of sex determination during
650 incubation of eggs of *Chelydra serpentina*. *J Morphol* 159: 17-27

651 Yntema CL Mrosovsky N (1980) Sexual differentiation in hatchling loggerheads
652 (*Caretta caretta*) incubated at different controlled temperatures.
653 *Herpetologica* 36: 33-36

654 Yntema CL Mrosovsky N (1982) Critical periods and pivotal temperatures for sexual
655 differentiation in loggerhead sea turtles. *Can J Zool* 60: 1012-1016
656
657
658

660 Table 1. Total nest and hatchling data for each state from 2000-2006 monitored beaches

State	Average annual number of nests	Total number of nests	Percent of total nests	Total number of hatchlings	Percent female	Percent male	Percent of total loggerhead population^a
Virginia	5	38	0.2%	1,366	24.9%	75.1%	0.01%
North Carolina	669	4,683	26.0%	347,812	48.9%	51.1%	1.03%
South Carolina	1,900	13,302	73.8%	867,763	57.0%	43.0%	2.92%
All	2,575	18,023	100%	1,216,941	53.5%	46.5%	3.96%

661 ^aBased on a mean estimated number of 65,040 loggerhead nests a year in the United States

662 from 2000-2006 (Florida Fish and Wildlife Conservation Commission 2007, M Dodd, pers.

663 comm).

664

665 FIGURE LEGENDS

666 Figure 1: Compared by Latitude a) Map showing range of study. Black dots indicate sites
667 from North to South: Assateague Island, VA; Sandbridge, VA; Back Bay National Wildlife
668 Sanctuary, VA; False Cape SP, VA; Northern Outerbank Beaches, NC; Pea Island National
669 Wildlife Refuge, NC; Cape Hatteras National Seashore, NC; Cape Lookout National
670 Seashore, NC; Fort Macon State Park, NC; Bogue Banks, NC; Bear Island, NC; Onslow
671 Beach, NC; Topsail Island, NC; Figure Eight Island, NC; Wrightsville Beach, NC; Pleasure
672 Island, NC; Fort Fisher State Park, NC; Long Beach/Oak Island, NC; Holden Beach, NC;
673 Ocean Isle, NC; Caswell Beach, NC; Sunset Beach/Bird Island, NC; Bald Head Island, NC;
674 City of Myrtle Beach, SC; Long Bay Estates, SC; Myrtle Beach State Park, SC; Horry
675 County, SC; Huntington Beach State Park, SC; Litchfield, SC; Pawleys Island, SC;
676 DeBordieu, SC; Hobcaw, SC; South Island, SC; Cape/Lighthouse Island, SC; Dewees Island,
677 SC; Isle of Palms/Sullivans Island, SC; Folly Beach, SC; Kiawah, SC; Seabrook, SC;
678 Edingsville, SC; Edisto Beach State Park, SC; Town of Edisto Beach, SC; Harbor Island, SC;
679 Hunting Island State Park, SC; Fripp Island, SC; Pritchards Island, SC; and Hilton Head, SC.
680 b) Mean annual number of nests per kilometer for each site; c) Mean incubation durations for
681 each site; d) Mean sex ratio for each site by latitude; e) Mean hatch success for each site from
682 2000-2006. Dotted lines indicate average for entire study site.

683

684 Figure 2: Range of incubation duration and sex ratios for study areas. a) and b) Virginia
685 1985-2006; c) and d) North Carolina 1995-2006; e) and f) South Carolina 2000-2006; g) and
686 h) All states 2000-2006. Dotted lines indicate median incubation durations/sex ratios.
687 Dashed lines indicate pivotal incubation which gives 50% of either sex. Sex ratio bars are
688 not continuous but come from incubation/sex ratio model and follow this pattern: 0%,
689 4.95%, 5.77%, 6.72%, 7.80%, 9.05%, 10.48%, 12.10%, 13.93%, 15.99%, 18.29%, 20.84%,
690 23.64%, 26.69%, 29.98%, 33.49%, 37.19%, 41.05%, 45.02%, 49.06%, 53.11%, 56.96%,
691 60.88%, 64.66%, 68.27%, 71.68%, 74.85%, 77.78%, 80.45%, 82.88%, 85.06%, 87.00%,

692 88.73%, 90.25%, 91.59%, 92.75%, 93.77%, 94.65%, 95.42%, 96.08%, 96.64%, 97.13%,
693 100%. Sex ratio patterns are clearly different for each state.

694

695 Figure3. Distribution of absorption of incident solar radiation by sand samples ($\alpha_{350-2500}$).

696 Average absorption was $56.76 \% \pm 5.47$.

697

698 Figure 4: Relationship between a) temperature and latitude b) incubation duration and latitude

699 c) temperature and sand absorption and d) incubation duration and sand absorption.

700 Temperature and sand absorption values were obtained in 2007 and incubation durations

701 were averaged for each site from 2000-2006. There are significant latitudinal trends for both

702 temperature and incubation. Deviations from these trends can be explained by differences in

703 sand absorption.

704

705 Figure 5: Annual sex ratios for six sites with long term incubation data: a) Back Bay National

706 Wildlife Refuge, VA ($36^{\circ}40' N 75^{\circ}54' W$); b) Cape Lookout National Seashore, NC ($34^{\circ}49'$

707 $N 76^{\circ}21' W$); c) Bald Head Island, NC ($33^{\circ}51' N 77^{\circ}59' W$); d) Cape/Lighthouse Island, SC;

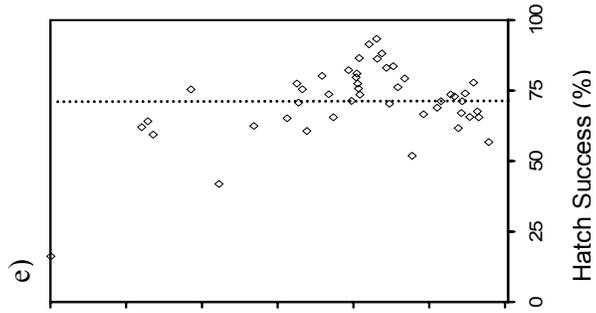
708 e) Edisto Beach State Park, SC ($32^{\circ}30' N 80^{\circ}18' W$); f) Hilton Head Island, SC ($32^{\circ}09' N$

709 $80^{\circ}43' W$). Dotted lines indicate average sex ratio. No significant long term trend in sex ratio

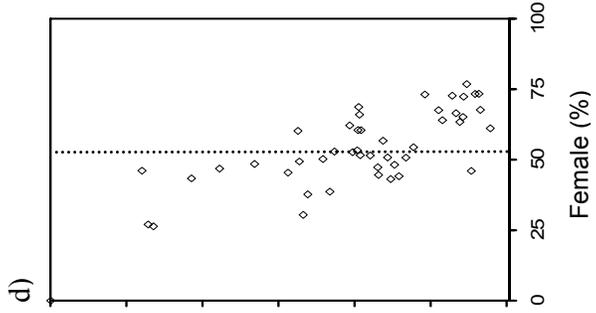
710 was found for any site except Bald Head Island (regression: $F_{1,24} = 17.6, p < 0.001, R^2 = 0.44$)

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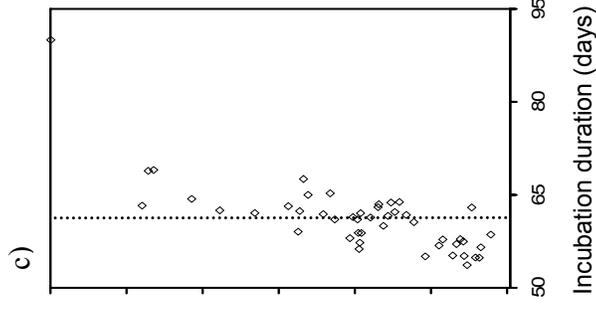
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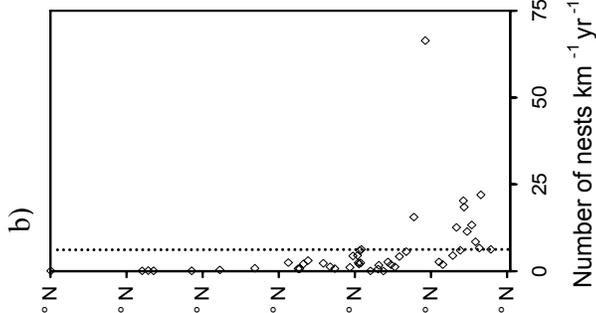
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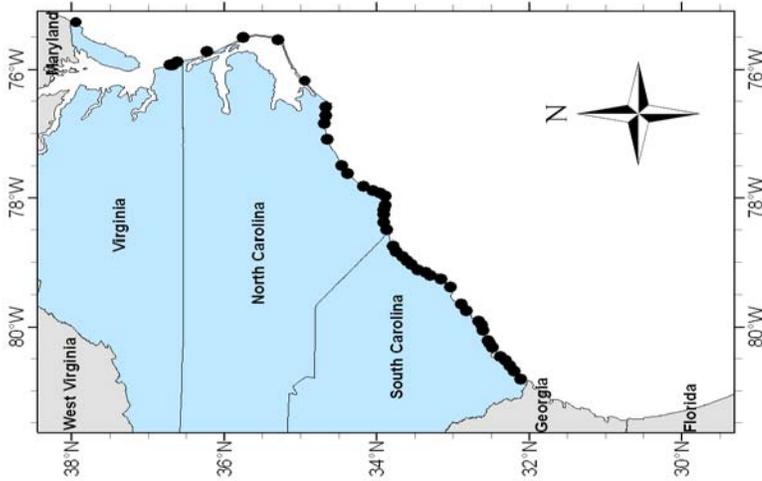


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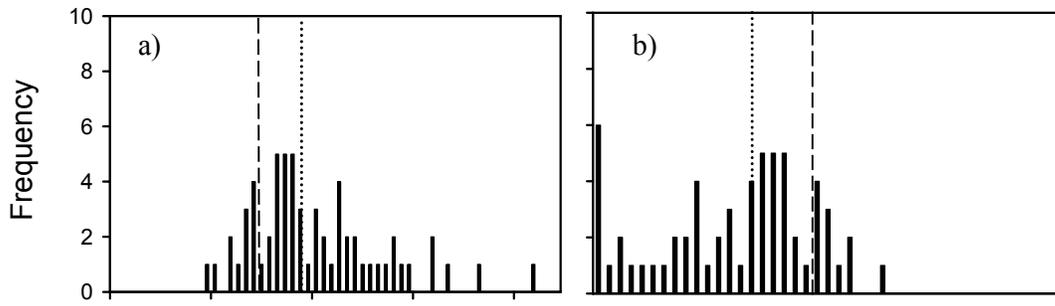


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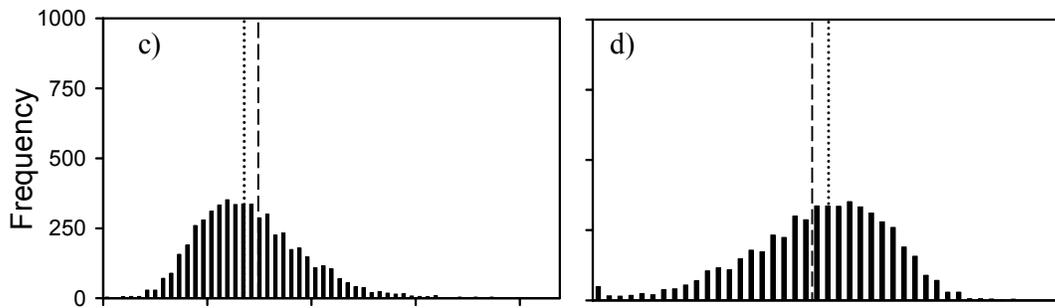
Figure 1. a)



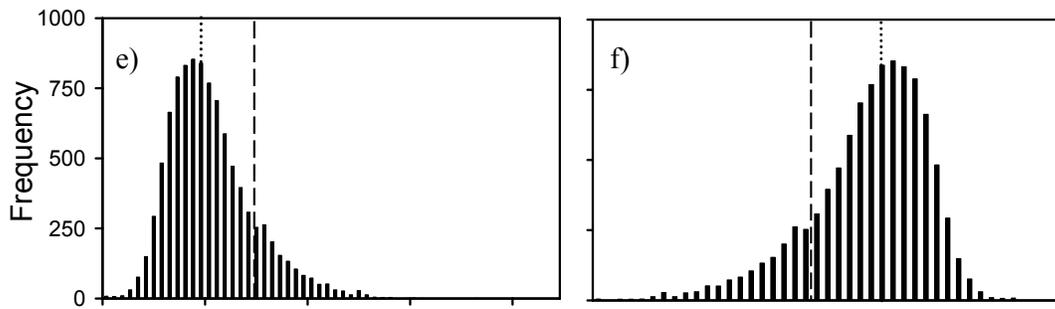
718 Figure 2.



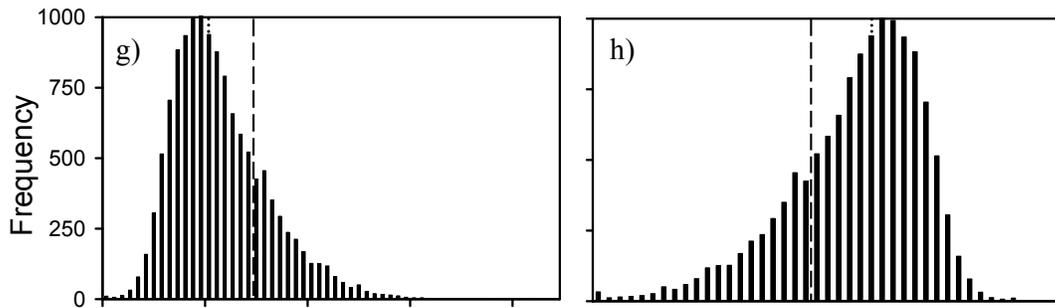
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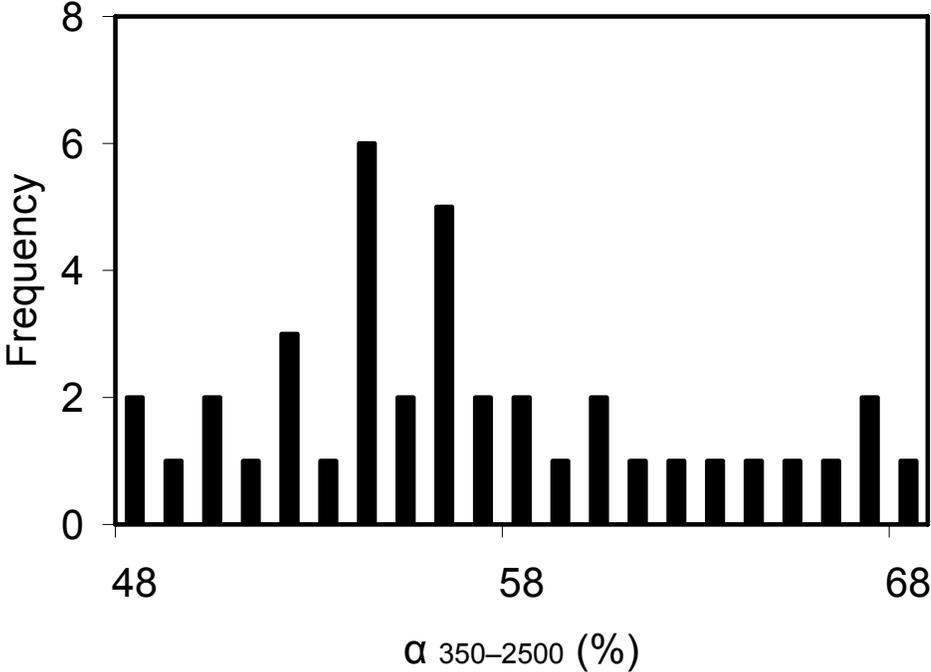
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Incubation duration (days)

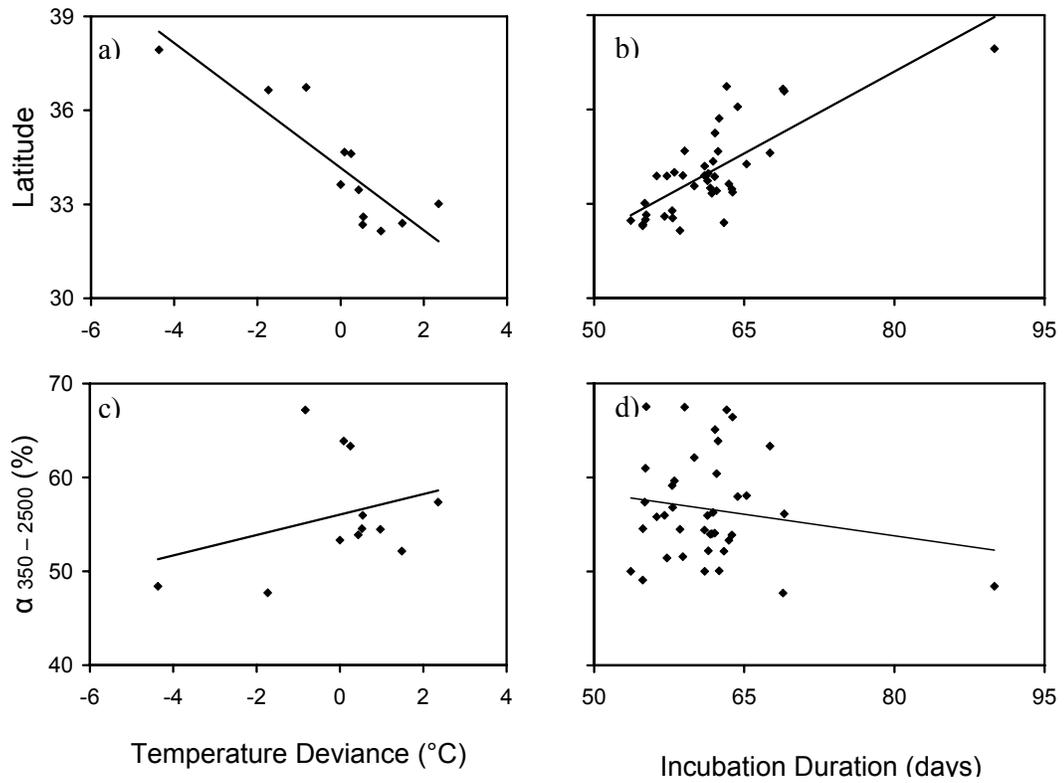
Percent female (%)

724 Figure 3.



725

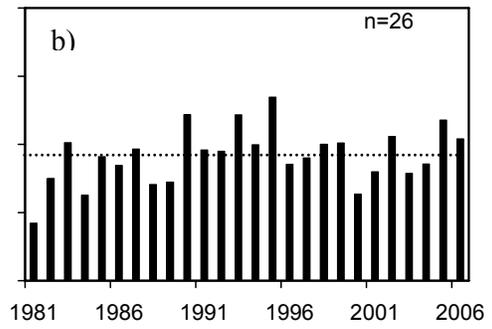
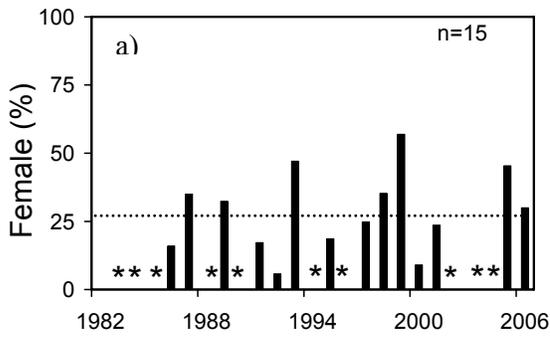
726 Figure 4.



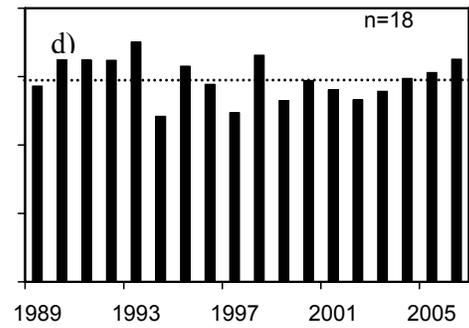
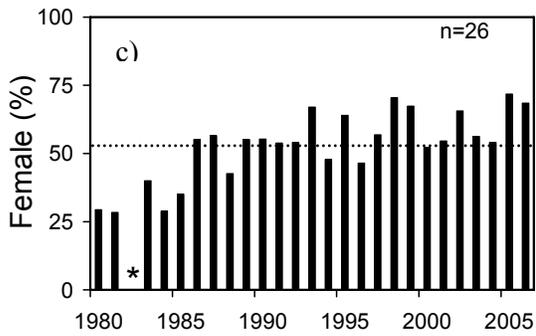
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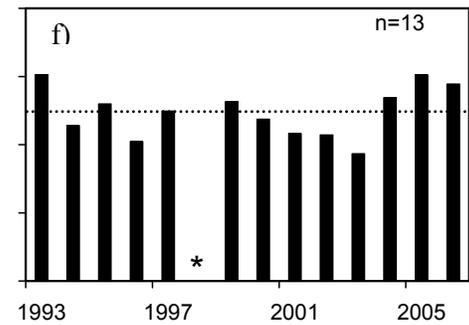
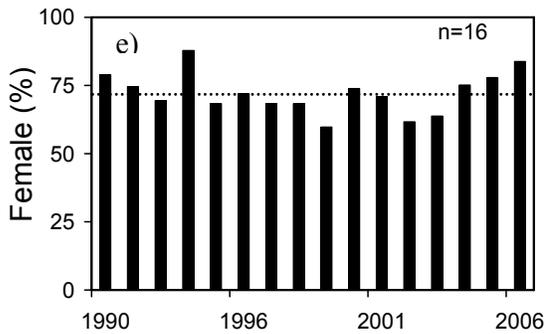
730 Figure 5.



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